

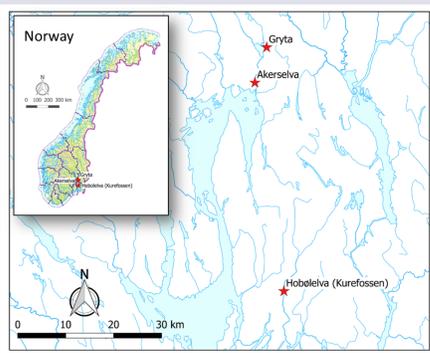
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Introduction

The ubiquitous pollution of various environments with microplastics (MP) is a global issue of growing concern. Freshwater environments, especially rivers, have gathered increasing attention as recent findings point out their role as relevant pathways of MP distribution. To see whether rivers with different characteristics carry different MP loads and polymer compositions, we investigated three rivers (city influenced, agricultural influenced and pristine) in South Eastern Norway regarding their MP concentration, loads, polymer composition and size class distribution.

Sampling



Sampling in three rivers of varying anthropogenic impact from low (green) to high (red) anthropogenic impact. Left to right: Gryta (pristine), Hobøl (agricultural influenced) and Akerselva (city influenced).

- Sampling of 1 m³ per river was conducted in May 2019
- A custom-built and well-established filtration device (UFO system)^[1] enabled sampling of MP in the range of 10–5000 μm
- Field blanks for airborne contamination were collected at each site

Sampling for MP (10–5000 μm)

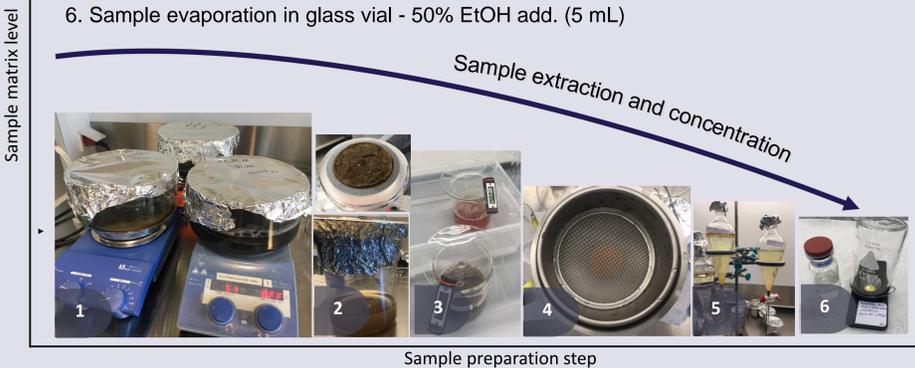


Custom-built filtration device (AAU UFO-system) with one 300 μm filter followed by two 10 μm filters installed

Sample preparation

Samples were processed through a multi-step enzymatic-oxidative sample treatment ^[2,3]:

1. SDS treatment (5% w.v.)
2. Two step enzymatic treatment (Protease; Cellulase & Viscozyme)
3. Fenton reaction (Fe (II) –catalysed)
4. Separation into two size fractions: 10–300 μm and >300 μm
5. Flotation with SPT (ρ = 1.7–1.8 g cm⁻³)
6. Sample evaporation in glass vial - 50% EtOH add. (5 mL)



Sample preparation step

ATR-FTIR & μFTIR-Imaging analysis & Pyr-GC/MS

Particles >300 μm:

Visual sorting of putative MP under stereo microscope



Analysis via Attenuated total reflectance Fourier-transform infrared spectroscopy (ATR-FTIR)



Particles 10–300 μm:

Deposition of a sub-sample on ZnSe window



Analysis via FPA-μFTIR-imaging and auto-detection via siMPle ^[4,5]



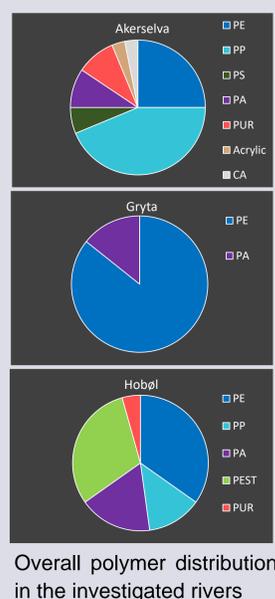
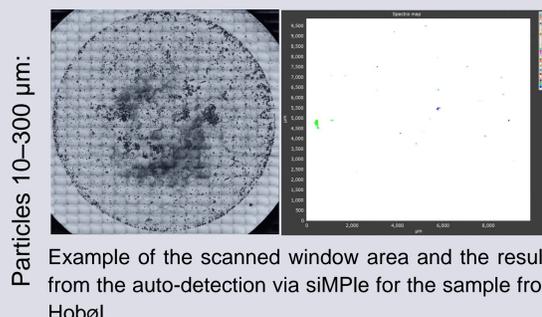
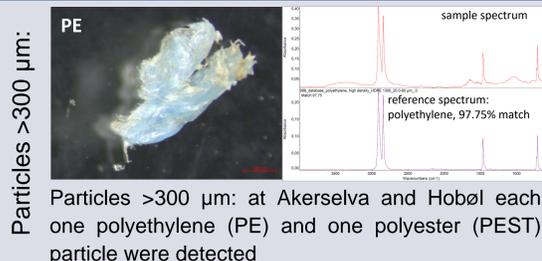
Filtration of sub-sample on GF/C filter



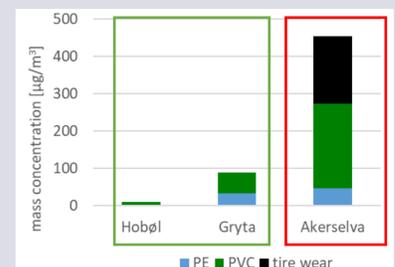
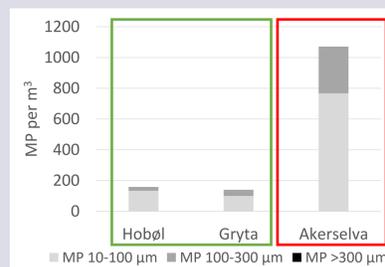
Analysis via pyrolysis gas chromatography/mass spectrometry (Pyr-GC/MS) ^[6]



Results & Discussion



The polymer type composition was different at the different sampling sites with the **city influenced** river Akerselva being the most diverse (6), followed by the **agricultural influenced** river Hobøl (5) and the **pristine** river Gryta (2). At Akerselva polypropylene (PP) was the most abundant polymer type, followed by polyethylene (PE), polyamide (PA), polystyrene (PS), polyurethane (PUR) and some other less common polymer types (e.g. acrylics). In Hobøl PE and polyester/polyethylene terephthalate (PEST) were most common followed by PA, PP and PUR. In Gryta only two polymers were detected, namely PE and PA.



MP concentrations and loads in the small size fraction (10–300 μm) were highest in the **city influenced** river Akerselva (1067 MP/m³ and 2740 μg/m³) followed by the two rivers in the **rural area**, Hobøl (153 MP/m³ and 50 μg/m³), and Gryta (140 MP/m³ and 450 μg/m³). Furthermore, The thermoanalytical analysis with Pyr-GC/MS revealed the presence of tire wear particles in the city influenced river.

Conclusion

Our study showed that the number of microplastic particles were 7 times larger in the city influenced river (Akerselva) compared with the two rivers from the rural areas (Hobøl and Gryta). The city influenced river also had the most diverse polymer type composition and was the only river to contain tire wear particles.

More than 70% of the MP were between 10 to 100 μm in size, followed by 13–28% in the 100–300 μm fraction and hardly any MP were larger than 300 μm. MP concentrations of the large size fraction (300–5000 μm) were 150–530 times lower than in the small size fraction. This highlights the importance of analysing the small size fraction (MP 10–300 μm) when considering the impact of land-based discharge of MP to the oceans.

References

- [1] Rist et al., 2020, <https://doi.org/10.1016/j.envpol.2020.115248>
- [2] Löder et al. 2017, <http://dx.doi.org/10.1021/acs.est.7b03055>
- [3] Liu et al. 2019, <https://doi.org/10.1016/j.scitotenv.2019.03.416>
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- [6] Gomiero et al. 2019, <https://doi.org/10.1016/j.chemosphere.2019.04.096>

